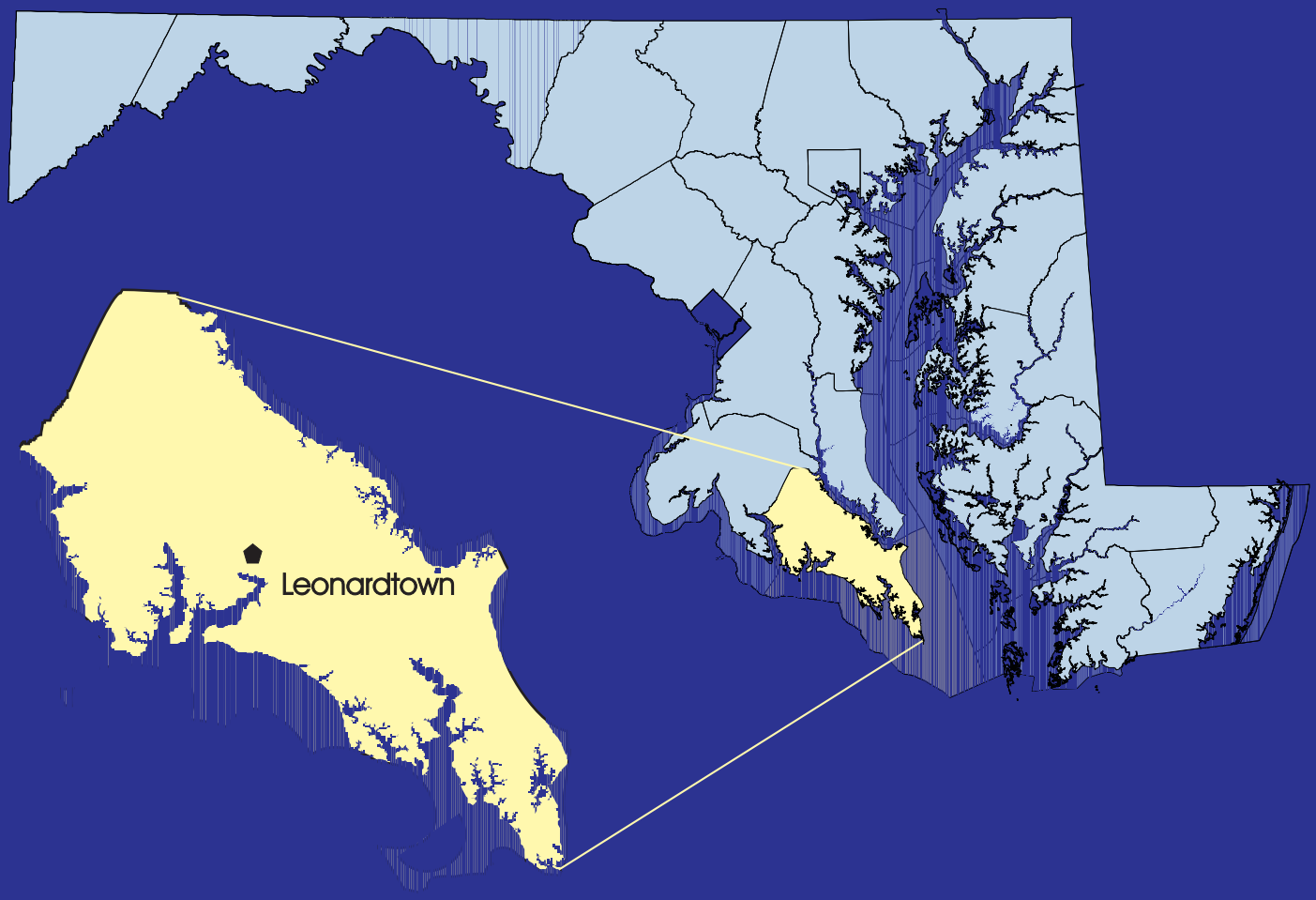
# ST. MARY'S COUNTY

## RESULTS OF THE 1994-1997 MARYLAND BIOLOGICAL STREAM SURVEY: COUNTY ASSESSMENTS




CHESAPEAKE BAY AND
WATERSHED PROGRAMS
MONITORING AND
NON-TIDAL ASSESSMENT

CBWP-MANTA-EA-01-24

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Publication date: December 2001
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# ST. MARY'S COUNTY

## Results of the 1994-1997 Maryland Biological Stream Survey: County-Level Assessments

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**December 2001**

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## FOREWORD

This report is based on results of the Maryland Biological Stream Survey (MBSS), a program funded primarily by the Power Plant Research Program and administered by the Maryland Department of Natural Resources (MDNR). Field data for the MBSS were collected by the Maryland Department of Natural Resources. Analyses of water chemistry samples were conducted by the University of Maryland's Appalachian Laboratory. Much of the initial data analysis was conducted by Versar, Inc. for MDNR's Power Plant Assessment Division.

This report helps fulfill two outcomes in MDNR's Strategic Plan: 1) A Vital and Life Sustaining Chesapeake Bay and Its Tributaries, and 2) Sustainable Populations of Living Resources and Healthy Ecosystems.

## ACKNOWLEDGMENTS

The 1994-1997 Maryland Biological Stream Survey has been a cooperative effort among several agencies, consultants and academic institutions. We wish to thank Nancy Roth and Ginny Mercurio from Versar in helping to compile some of the data used in this report. Versar also designed the sampling program, obtained landowners' permissions, and helped manage the data. We are also grateful to the many individuals from Maryland Department of Natural Resources, the University of Maryland's Appalachian Laboratory (AL), and the University of Maryland's Wye Research and Education Center (WREC) who comprised the field crews and did a great job collecting the data. MDNR staff also digitized watersheds and calculated land use data, provided quality assurance, and conducted field crew training. Nancy Roth and her colleagues at Versar developed the fish Index of Biotic Integrity, and Dr. Sam Stribling and his staff at Tetra Tech, Inc. developed the benthic Index of Biotic Integrity. Dr. Ray Morgan of AL and Mr. Lenwood Hall of the WREC supervised additional field crews and developed the Physical Habitat Index, and Dr. Keith Eshleman of AL assisted with analyses of data on acidified streams. Drs. Wayne Starnes and Bob Reynolds of the Smithsonian Institution (reptiles and amphibians), Dr. Rich Raesly of Frostburg State University (fish), Rita Villella of the U.S. Geological Survey Leetown Science Center (mussels), and Michael Naylor of MDNR (aquatic vegetation) provided taxonomic verifications of voucher specimens. The success of the project resulted from the strong efforts of all these groups. Special thanks go to Ron Klauda for his editorial support and Brenda Morgan for her assistance in formatting, editing, and organizing the report.

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## INTRODUCTION

This report presents county-level data from the 1994-1997 Maryland Biological Stream Survey (MBSS or the Survey). Previous reports have documented interim results from the 1995 (Roth et al. 1997) and 1996 (Roth et al. 1998a) sample years. In addition, a comprehensive final report was produced to assess the "state of the streams" throughout the state (Roth et al. 1999). All previous MBSS reports have presented information by individual drainage basins. Because there is a recognized need for stream health information at the county level, a series of reports were prepared; this report is part of that series. This introductory section recounts the origin of the Survey and describes its components.

### *Origin of the MBSS*

More than 10 years ago, the Maryland Department of Natural Resources (MDNR) recognized that atmospheric deposition was one of the most important environmental problems resulting from the generation of electric power. To determine the extent of acidification of Maryland streams resulting from acidic deposition, MDNR conducted the Maryland Synoptic Stream Chemistry Survey (MSSCS) in 1987. The MSSCS estimated the number and extent of streams at that time affected by or sensitive to acidification statewide and demonstrated the potential for adverse effects on biota from acidification. However, little direct information was available on the biological responses of Maryland streams to water chemistry conditions. Data that were available could not be used (because of methodological differences and spatial coverage limitations) to compare conditions across regions or watersheds (Tornatore et al. 1992). Neither was it possible to assess the interactions between acidic deposition and other anthropogenic and natural influences (CBRM 1989). For these reasons, in 1993, MDNR created the MBSS to provide comprehensive information on the status of biological resources in Maryland streams and how they are affected by acidic deposition and other cumulative effects of anthropogenic stresses.

### *Description of the MBSS*

The MBSS is intended to help environmental decision-makers protect and restore the natural resources of Maryland. The primary objectives of the MBSS are:

- to assess the current status of biological resources in Maryland's non-tidal streams;
- to quantify the extent to which acidic deposition has affected or may be affecting biological resources in the state;
- to examine which other water chemistry, physical habitat, and land use factors are important in explaining the current status of biological resources in streams;
- to compile the first statewide inventory of stream biota;
- to establish a benchmark for long-term monitoring of trends in these biological resources; and
- to target future local-scale assessments and mitigation measures needed to restore degraded biological resources.

In creating the Survey, MDNR implemented a probability-based sampling design as a cost-effective way to characterize statewide stream resources. By randomly selecting sites, the Survey can make quantitative inferences about the characteristics of all 9,258 miles of first-to-third-order, non-tidal streams in Maryland (based on stream length on a 1:250,000-scale base map). MDNR recognized that the utility of these estimates depended on accurately measuring appropriate attributes of streams. The Survey focuses on biology for two reasons: (1) organisms themselves have direct societal value and (2) biological communities integrate stresses over time and are a valuable and cost-effective means of assessing ecological integrity (i.e., the capacity of a resource to sustain its inherent potential).

Fish are an important component of stream integrity and one that also contributes to substantial recreational values. For these reasons, fish communities are a primary focus of the Survey. The Survey collects quantitative data for the calculation of population estimates for individual fish species (both game and nongame). These data can also be used to evaluate

fish community composition, individual fish health, and the geographic distribution of commercially important, rare, or non-indigenous fish species. Benthic (bottom-dwelling) macroinvertebrates are another essential component of streams and they constitute the second principal focus of the Survey. The Survey uses rapid bioassessment procedures for collecting benthic macroinvertebrates; these semi-quantitative methods permit comparisons of relative abundance and community composition, and have proven to be an effective way of assessing biological integrity in streams (Hilsenhoff 1987, Lenat 1988, Plafkin et al. 1989, Kerans and Karr 1994, Resh 1995). The Survey also records the presence of reptiles and amphibians (herpetofauna), freshwater mussels, and aquatic plants (both submerged aquatic vegetation (SAV) and emergent macrophytes). The Survey has established rigorous protocols (Kazyak 1996) for each of these sampling components, as well as training and auditing procedures to assure that data quality objectives are met.

Although the MBSS sampling design and protocols provide exceptional information for characterizing the stream resources in Maryland, designation of degraded areas and identification of likely stresses requires additional activities. Assessing the condition of biological resources (whether they are degraded or not degraded) requires the development of ecological indicators that permit the comparison of sampled segment results to minimally impacted reference conditions (i.e., the biological community expected in watersheds with little or no human-induced impacts). The Survey has used its growing database of information collected with consistent methods and broad coverage across the state to develop and test indicators of individual biological components (Stribling et al. 1998, Roth et al. 1998b) and physical habitat quality (Hall et al. 1999). Each of these indicators consists of multiple metrics using the general approach developed for the Index of Biotic Integrity (IBI) (Karr et al. 1986, Karr 1991) and the Chesapeake Bay Benthic Restoration Goals (Ranasinghe et al. 1994). The fish and benthic macroinvertebrate IBIs (which combine attributes of both the number and the type of species found) are widely accepted indicators that have been adapted for use in a variety of geographic locations (Miller et al. 1988, Cairns and Pratt 1993, Simon 1999). The Survey is investigating the possibility of developing additional indicators (e.g., amphibians in small streams with few or no fish) and combining components into a composite indicator of biological integrity.

In addition to developing reference-based indicators, the Survey is applying a variety of analytical methods to the question of which stressors are most closely associated with degraded streams. This involves correlational and multivariate analyses of water chemistry, physical habitat, land use, and biological information (e.g., presence of non-native species). The biological information also provides a valuable opportunity for documenting aquatic biodiversity across the state; the distribution and abundance of species previously designated as rare only by anecdotal evidence can be determined, and unique combinations of species at the ecosystem and landscape levels can be identified. Land use and other landscape-scale metrics will play an important role in identifying the relative contributions of different stressors to the cumulative impact on stream resources. Ultimately, the Survey seeks to provide an integrated assessment of the problems facing Maryland streams that will facilitate interdisciplinary solutions for their restoration. The survey also provides resource managers with the locations of relatively undisturbed streams and watersheds that deserve protection.

## METHODS

This section presents the specific study design and procedures used to implement the Maryland Biological Stream Survey. The study area of concern and the sampling design developed to characterize it are presented, along with field and laboratory methods for each component: fish, benthic macroinvertebrates, reptiles and amphibians, physical habitat, and water chemistry. Methods for aquatic vegetation and mussel sampling are presented, but the resulting data are not included in this report. A full description of MBSS methods can be found in Kazyak (1996).

### MBSS Study Design

The Survey study area comprises 17 distinct drainage basins across the state. Random sampling was used to allow the estimation of unbiased summary statistics (e.g., means, proportions, and their respective variances) for the entire state, a particular basin, and subpopulations of interest (e.g., streams with pH < 5).

Because it would have been cost prohibitive to visit a sufficient number of sites in all basins in a single year, lattice sampling was used to schedule sampling of all basins over a three-year period, 1995-1997. Lattice sampling, also known as multistratification, is a cost-effective means of allocating effort across time in a large geographic area (Heimbuch 1999, Jessen 1978, Cochran 1977). A table, or lattice, was formed by arranging 17 basins in 17 rows, and the years in 3 columns. Lattice sampling was the method used for selecting cells from this 17x3 table so that all basins would be sampled over a three-year period and all basins would have a non-zero probability of being sampled in a given year. The data presented in this report include those collected at random sampling sites within the 17 principal basins in Maryland, as well as sites from the 1994 demonstration project. Because no estimates were calculated for this report, these data were included to supplement the number of sites.

The sampling frame for the Survey was constructed by overlaying basin boundaries on a map of all blue-line stream reaches in the study area as digitized on a U.S. Geological Survey 1:250,000 scale topographic map. This sample frame was similar to that used by the earlier Maryland Synoptic Stream Chemistry Survey (MSSCS) conducted in 1987 (Knapp and Saunders 1987, Knapp et al. 1988). The Strahler convention (Strahler 1957) was used for ranking stream reaches by order; first-order reaches, for example, are the most upstream reaches in the branching stream system. Sampling was restricted to non-tidal, third-order and smaller stream reaches, excluding impoundments that were non-wadable or that substantially altered the riverine nature of the reach (Kazyak 1994). Together, these first-through third-order streams comprise about 90% of all stream and river miles in Maryland. Stream reaches were further divided into non-overlapping, 75-meter segments; these segments were the elementary sampling units from which biological, water chemistry, and physical habitat data were collected.

The 1995-1997 MBSS study design was based on stratified random sampling of segments within each basin; each basin was stratified by stream order. Within a stream order, the number of segments sampled per basin is proportional to the number of stream miles in the basin. To achieve the target number of samples per stream order within each basin, a given number of segments were randomly selected from each basin and ranked in order of selection. In all basins, extra segments were selected as a contingency against loss of sampling sites from restricted access to selected streams or from streams that were dry, too deep, or otherwise unsampleable owing to field conditions. In some basins, where only a small number of sites would have been selected using this method, additional random sites were selected to increase sample size. These extra sites (selected at random using the method described above) were used to provide better basinwide estimates; they were not included in the estimates of statewide conditions.

Permissions were obtained to access privately owned land adjacent to or near each stream segment. The procedures for obtaining permissions are described in Chaillou (1995). Because landowner permissions were obtained in a synoptic fashion and some variation in these rates occurred, we obtained more permissions than were needed for the Survey. Only the highest ranking sites were sampled until the target goal for that basin was reached. For the three year study, the success rate for obtaining permission to access stream sampling segments was high. Eighty-eight percent of sites that were targeted for permission were sampled.

Reasons for permission denial varied and generally reflected the preferences of landowners regarding property access, rather than any specific types of land. In rare cases, permission denial may affect the interpretation of Survey estimates, but only where denials occur in streams with characteristics that differ from the general population of streams. In one example of potential bias, several sites with known coal mining activities in the North Branch Potomac basin denied permission to sample, likely under representing the proportion of acid mine drainage streams in the population.

***Field and Laboratory Methods***

Benthic macroinvertebrate and water quality sampling were conducted in spring, when the benthos are thought to be reliable indicators of environmental stress (Plafkin et al. 1989) and when acid deposition effects are often the most pronounced. Fish, reptiles and amphibians, aquatic vegetation, and mussel sampling, along with physical habitat evaluations, were conducted during the low-flow period in summer. Fish community composition tends to be stable during summer, and low flow is advantageous for electrofishing. Because low-flow conditions in summer may be a primary factor limiting the abundance and distribution of fish populations, habitat assessments were performed during the summer. The sample size in summer is lower than in spring because some streams were dry in summer or were, in rare cases, otherwise unsampleable.

To reduce temporal variability, sampling during spring and summer was conducted within specific, relatively narrow time intervals, referred to as index periods (Janicki et al. 1993). These index periods were defined by degree-day limits for specific parts of the state. This approach provided a synoptic assessment of the current status of stream biota, water quality, and physical habitat in the 17 basins sampled. The spring index period was the time period between approximately March 1 and May 1, with end of the index period determined by degree-day accumulation as specified in Hilsenhoff (1987). In reality, most spring samples (78%) were collected in March, well before degree-day accumulation limits were approached. The summer index period was between June 1 and September 30 (Kazyak 1994).

***Data Collection and Measurement***

Field sampling followed procedures specified in the MBSS sampling manual (e.g., Kazyak 1996). A summary of the variables measured and the field and laboratory methods used to conduct the sampling follows.

***Fish***

Fish were sampled during the summer index period using double-pass electrofishing within 75-meter stream segments. Block nets were placed at each end of the segment and direct current backpack electrofishing units were used to sample the entire segment. An attempt was made to thoroughly fish each segment, and consistent effort was applied over the two passes. This sampling approach allowed calculation of several metrics useful in calculating a biological index and produced unbiased estimates of fish species abundance.

In small streams, a single electrofishing unit was used. In larger streams, two to five units were employed to effectively sample the site. Captured fish were identified to species, counted, weighed, and released. Any individuals that could not be identified to species were retained for laboratory confirmation. For each pass, all individuals of each gamefish species (defined as trout, bass, walleye, pike, chain pickerel, and striped bass) were measured for total length and examined for visible external pathologies or anomalies. For nongame species, up to 100 fish of each species (from both passes) were examined for visible external pathologies or anomalies. For each pass, all non-game species were weighed together for an aggregate biomass measurement; gamefish were also weighed in aggregate to the nearest 10 g.

Electrofishing was also conducted at supplemental, non-randomly selected sites during the summer index period. The presence of each species of fish was recorded for these segments to provide additional qualitative information on statewide fish distributions. Sampling effort at most qualitative sites was based on doubling the elapsed time since the last species was recorded or a minimum of 600 seconds of electrofishing effort.

After processing the fish collected in the field, voucher

specimens were retained for each species not previously collected in the drainage basin. In addition, all individuals which could not be positively identified in the field were retained. The remaining fish were released. All voucher specimens and fish retained for positive identification in the laboratory were examined and verified by the MBSS Quality Assurance Officer or ichthyologists at Frostburg State University, Frostburg, Maryland or the Smithsonian Institution, Washington, DC.

***Benthic Macroinvertebrates***

Benthic macroinvertebrates were collected to provide a qualitative description of the community composition at each sampling site (Kazyak 1996). Sampling was conducted during the spring index period. Benthic community data were collected for the purpose of calculating biological metrics, such as those described in EPA's Rapid Bioassessment Protocols (Plafkin et al. 1989), and use as an indicator of biological integrity for Maryland streams.

At each segment, a 600 micron mesh "D" net was used to collect organisms from habitats likely to support the greatest taxonomic diversity. A riffle area was preferred, but other habitats were also sampled using a variety of techniques including kicking, jabbing, and gently rubbing hard surfaces by hand to dislodge organisms. If available, other habitat types were sampled, including rootwads, woody debris, leaf packs, macrophytes, and undercut banks. Each jab covered one square foot, and a total of approximately 2.0 $m^2$ (20 square feet) of combined substrates was sampled and preserved in 70% ethanol. In the laboratory, the preserved sample was transferred to a gridded pan and organisms were picked from randomly selected grid cells until the cell that contained the 100th individual (if possible) was completely picked. Some samples had fewer than 100 individuals. The benthic macroinvertebrates were identified to genus, or lowest practicable taxon, in the laboratory.

***Index of Biotic Integrity***

Sites were evaluated using both the fish (F-IBI) and benthic macroinvertebrate (B-IBI) IBIs developed for the MBSS (for detailed methods, see Roth et al. 1997 and Stribling et al. 1998). IBI scores for the MBSS are determined by comparing the fish or benthic macroinvertebrate assemblages at each site to those found at minimally impacted reference sites. Three separate formulations were employed for the fish IBI, one for each of three distinct geographic areas: Coastal Plain, Eastern Piedmont, and Highland. The two formulations used for the benthic IBI cover the Coastal Plain and non-Coastal Plain regions. Individual metrics for the IBI are scored 1, 3, or 5, based on comparison with the distribution of metric values at reference sites. For either the individual metrics or total IBI, a score of 3 or greater is considered comparable to reference site conditions, while scores falling below this threshold differ significantly from the reference conditions. Scores for the MBSS IBIs are calculated as the mean of the individual metric scores and therefore range from 1 to 5. Some other programs have used a similar approach (e.g., Weisberg et al. 1997), while others have instead computed the IBI as the total of individual metric scores. For example, Karr et al. (1986) calculated IBI as the sum of 12 metric scores, with totals ranging from 12 to 60 points.

***Reptiles and Amphibians***

At each sample segment, reptiles and amphibians were identified and the presence of observed species was recorded during the summer index period. A search of the riparian area was conducted within 5 meters of the stream on both sides of the 75-meter segment. Any reptiles and amphibians collected during the electrofishing of the stream segment were also included in the species list. Individuals were identified to species when possible. Voucher specimens and individuals not positively identifiable in the field were retained for examination in the laboratory and confirmation by herpetologists at the Smithsonian Institution, Washington, DC, or Towson University, Towson, Maryland.

***Physical Habitat***

Habitat assessments were conducted at all stream segments as a means of assessing the importance of physical habitat to the biological integrity and fishability of freshwater streams in Maryland. Procedures for habitat assessments (Kazyak 1996) were derived from two currently used methodologies: EPA's Rapid

Bioassessment Protocols (RBPs) (Plafkin et al. 1989), as modified by Barbour and Stribling (1991), and the Ohio EPA's Qualitative Habitat Evaluation Index (QHEI) (Ohio EPA 1987, Rankin 1989). A number of characteristics (instream habitat, epifaunal substrate, velocity/depth diversity, pool/glide/eddy quality, riffle/run quality, channel alteration, bank stability, embeddedness, channel flow status, and shading) were assessed qualitatively, based on visual observations within each 75-meter sample segment. Riparian zone vegetation width was estimated to the nearest meter, up to 50 meters from the stream. Additional observations of the surrounding area were used to assign ratings for aesthetic value (based on visible signs of human refuse at a site) and remoteness (based on distance from the nearest road, accessibility, and evidence of human activity). Also recorded were the presence or absence of various stream features including substrate types, various morphological characteristics, beaver ponds, point sources, and stream channelization. Local land uses visible from the stream segment and riparian vegetation type were also noted. Several additional physical characteristics were measured quantitatively to further characterize the habitat for each segment (see Kazyak 1996 for details). Quantitative measurements of the segment included maximum depth, stream gradient, velocity, thalweg depth, number of functional rootwads, number of functional large woody debris, wetted width, sinuosity, and overbank flood height. A velocity/depth profile was measured or other data were collected to enable calculation of discharge.

***Physical Habitat Index***

The Physical Habitat Index (PHI) was developed using MBSS data from 1994 to 1997 (Hall et al. 1999). As was the case in development of the fish and benthic IBIs, the conceptual approach was based on evaluating the relative importance (discriminatory power) of individual metrics and combinations of metrics explaining natural differences in streams throughout Maryland. These metrics were derived from both quantitative and qualitative habitat data collected during the summer index period. Based on analyses conducted for both fish IBI (Roth et al. 1998) and benthic macroinvertebrate IBI (Stribling et al. 1998) development in Maryland, the State was divided into two regions: the Coastal Plain and non-Coastal Plain. The resulting index was then adjusted to a centile scale that rated each sample segment as follows: Good - 72 to 100; Fair - 42 to 71.9; Poor - 12 to 41.9; and Very Poor - 0 to 11.9.

***Water Chemistry***

During the spring index period, water samples were collected at each site for analysis of pH, acid neutralizing capacity (ANC), conductivity, sulfate, nitrate-nitrogen, and dissolved organic carbon (DOC). These variables describe basic water quality conditions with an emphasis on factors related to acidic deposition.

Grab samples were collected in one-liter bottles for analysis of all analytes except pH. Water samples for pH were collected with 60 ml syringes, which allowed purging of air bubbles to minimize changes in carbon dioxide content (EPA 1987). Samples were stored on wet ice and shipped on wet ice to the analytical laboratory within 48 hours. Laboratory analyses were carried out by the University of Maryland's Appalachian Laboratory in Frostburg.

Chemical analysis of water samples followed standard methods described in EPA's Handbook of Methods for Acid Deposition Studies (EPA 1987). EPA protocols were followed, except that ANC sample volume was reduced to 40 ml to ease handling. Routine daily quality control (QC) checks included processing duplicate, blank, and calibration samples according to EPA guidelines for each analyte. Field duplicates were taken at 5% of all sites. Routine QC checks helped to identify and correct errors in sampling routines or instrumentation at the earliest possible stage.

During the summer index period, in situ measurements of dissolved oxygen (DO), pH, temperature, and conductivity were collected at each site to further characterize existing water quality conditions that might influence biological communities. Measurements were made at an undisturbed section of the segment, usually in the middle of the stream channel, using electrode probes. Instruments were calibrated daily and calibration logbooks were maintained to document instrument performance.

Recognizing that water temperature is an important factor affecting stream condition, but one that varies

daily and seasonally, temperature loggers were deployed at 220 sites in five basins during 1997. The basins sampled were: the Choptank, Susquehanna, Potomac Washington Metro, Patuxent, and Pocomoke. Onset Computer Corporation Optic Stowaway temperature loggers were anchored in each site during the summer index period. Water temperature was recorded every 15 minutes from June 15 until mid-September.

***Mussels***

During the summer index period, freshwater mussels were sampled qualitatively by examining each 75-meter stream segment for their presence. Mussels were identified to species, their presence recorded, and subsequently released. Species not positively identifiable in the field were retained for confirmation by U.S. Geological Survey (USGS) Biological Resources Division staff.

***Aquatic Vegetation***

Aquatic vegetation was sampled qualitatively by examining each 75-meter segment for the presence of aquatic plants. Plants were identified to species and their presence recorded for each site. While the primary objective was to document the presence of submerged aquatic vegetation (SAV), emergent and floating aquatic vegetation was also recorded when encountered. Species not positively identifiable in the field were retained for laboratory examination and confirmation by MDNR's staff expert on SAV. Due to the difficulty in long-term preservation, no permanent vouchers of aquatic vegetation were retained.

***Data Management***

All crews used standardized pre-printed data forms developed for the Survey to ensure that all data for each sampling segment were recorded and standard units of measure were used (Kazyak 1996). Using standard data forms facilitated data entry and minimized transcription error. The field crew leader and a second reviewer checked all data sheets for completeness and legibility before leaving each sampling location. Original data sheets were sent to the Data Management Officer for further review and data entry, while copies were retained by the field crews.

A custom database application, in which the input module was designed to match each of the field data sheets, was used for data entry. Data were independently entered into two databases and compared using a computer program as a quality-control procedure. Differences between the two databases were resolved from original data sheets or through discussions with field crew leaders.

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# Maryland Biological Stream Survey Data

## COUNTY SUMMARY

A total of 18 quantitative sites were sampled in St. Mary's County by MBSS sampling crews during 1994-1997 (Table 1; Figure 2). Qualitative fish sampling was conducted at an additional nine sites to provide a more complete picture of fish species distributions. Appendix A provides a summary of the types of data available for each of the sites sampled.

### Species Highlights

Only 19 fish species were collected (Table 2). This number ties for a ranking of next to last among the counties in the state. The most commonly found fish species were eastern mudminnow and American eel. No rare fish species were collected in St. Mary's County during 1994-1997, but a population of ironcolor shiner was discovered by an MBSS crew during sampling in 2000. Ironcolor shiner is a species currently being considered for listing as threatened or endangered in Maryland.

The 131 genera of benthic macroinvertebrates found ranks seventeenth among Maryland counties for benthos diversity (Table 3). Forty-eight of those genera were collected at a single site. Only 4% of the taxa could be found at more than half of the sites sampled.

Fifteen species of reptiles and amphibians were found in or near St. Mary's County streams (Table 4), tying the county for a ranking of fifteenth in the state. No state or federally listed reptiles or amphibians were collected during the sampling.

### Ecological Health

Based on the indicators currently in use by the MBSS, the overall ecological health of St. Mary's County's headwater streams in the county can best be described as Fair. The average F-IBI score was 3.35 (lower end of the Fair range, tying for a ranking of tenth among counties in the state) and the average B-IBI score among sites was 3.2 (low end of the Fair category, but ranking second best among counties in the state). Based on F-IBI and B-IBI scores from individual sites, the highest rated stream is Burnt Mill Branch (Table 6). The worst rated streams include Cecil Creek and Penbrook Run. It should be noted that a modified IBI for fish in acidic, blackwater streams is being developed, and that F-IBI scores could presumably improve when the blackwater IBI is applied to streams in the county.

### Physical Habitat

Physical habitat in St. Mary's County was rated as Fair by the Physical Habitat Index. Values ranged from 4.74 to 94.1, with an average score of 52.10 (low end of the Fair range, ranking thirteenth among counties in the state) (Table 6; Figure 5). Other noteworthy points about county streams include a ranking of first in the state for riparian buffer width and shading, a ranking of second best for the amount of trash found in streams, and rankings of fourth and sixth best for instream rootwads (trees whose roots protect banks from erosion and provide habitat for aquatic life) and large woody debris (instream logs), respectively. Both rootwads and woody debris are crucial to healthy stream systems, especially in streams dominated by sand and silt.

### Nitrate-Nitrogen

Nitrate-nitrogen values at sites sampled averaged 0.63 mg/L, or fourth best in the state. The stream with the lowest nitrate value was St. Thomas Creek, and the stream with the highest nitrate value was an unnamed tributary to Killpeck Creek (Table 7). In no stream was the EPA limit for drinking water (10 mg/L) exceeded.

**Table 1.** Site information and land use data collected at Maryland Biological Stream Survey sites in St. Mary's County, 1994-1997. Basin abbreviations are as follows: LP - Lower Potomac River; PX - Patuxent River.

| Site | Latitude | Longitude | Stream Name | Basin | Order | Catchment Acres | % Urban | % Agric. | % Forest |
|---|---|---|---|---|---|---|---|---|---|
| SM-S-006-212-95 | 38.3477 | 76.6452 | Burnt Mill Creek | LP | 2 | 3402.07 | 3.56 | 34.89 | 54.96 |
| SM-S-007-138-95 | 38.2512 | 76.4443 | Penbrook Run | LP | 1 | 453.41 | 27.78 | 12.31 | 55.93 |
| SM-S-036-107-95 | 38.3248 | 76.7477 | Dynard Run | LP | 1 | 903.81 | 0.19 | 38.09 | 61.16 |
| SM-S-039-127-95 | 38.3127 | 76.7058 | Ut St Clements Bay | LP | 1 | 569.69 | 1.79 | 61.17 | 37.04 |
| SM-S-040-128-95 | 38.3730 | 76.6635 | Ut Burnt Mill Creek | LP | 1 | 366.37 | 8.97 | 23.18 | 63.27 |
| SM-S-051-132-95 | 38.3666 | 76.7372 | Ut St Clements Creek | LP | 1 | 215.91 | 0.00 | 20.75 | 75.47 |
| SM-S-090-238-97 | 38.4810 | 76.7690 | Un Trib To Killpeck Cr | PX | 2 | 422.65 | 29.17 | 25.14 | 45.69 |
| SM-S-104-126-95 | 38.3878 | 76.6782 | Ut St Clements Creek | LP | 1 | 125.25 | 26.97 | 20.70 | 44.42 |
| SM-S-111-112-95 | 38.2566 | 76.4998 | Jarboesville Run | LP | 1 | 2025.90 | 15.80 | 22.78 | 52.12 |
| SM-S-116-214-95 | 38.3916 | 76.8538 | Budds Creek | LP | 2 | 3188.51 | 3.03 | 21.73 | 73.57 |
| SM-S-125-142-97 | 38.3410 | 76.5540 | Cuckold Cr | PX | 1 | 223.68 | 23.59 | 18.42 | 57.84 |
| SM-S-166-143-97 | 38.3770 | 76.5820 | Un Trib To Coles Cr | PX | 1 | 91.60 | 10.00 | 28.51 | 61.48 |
| SM-S-199-302-95 | 38.3484 | 76.7303 | St Clements Creek | LP | 3 | 10875.26 | 6.63 | 36.23 | 52.60 |
| SM-S-209-105-95 | 38.2851 | 76.6931 | Cecil Creek | LP | 1 | 590.37 | 0.00 | 30.74 | 65.45 |
| SM-S-214-1-94 | 38.3669 | 76.5817 | St Thomas Cr | PX | 1 | 144.20 | 28.29 | 7.77 | 63.73 |
| SM-S-214-2-94 | 38.3681 | 76.5735 | St Thomas Cr | PX | 1 | 422.20 | 10.80 | 12.32 | 74.70 |
| SM-S-237-103-95 | 38.4460 | 76.7688 | Chaptico Creek | LP | 1 | 772.14 | 19.73 | 39.47 | 33.57 |
| SM-S-239-310-95 | 38.3402 | 76.7281 | St Clements Creek | LP | 3 | 11608.80 | 6.35 | 37.62 | 51.27 |

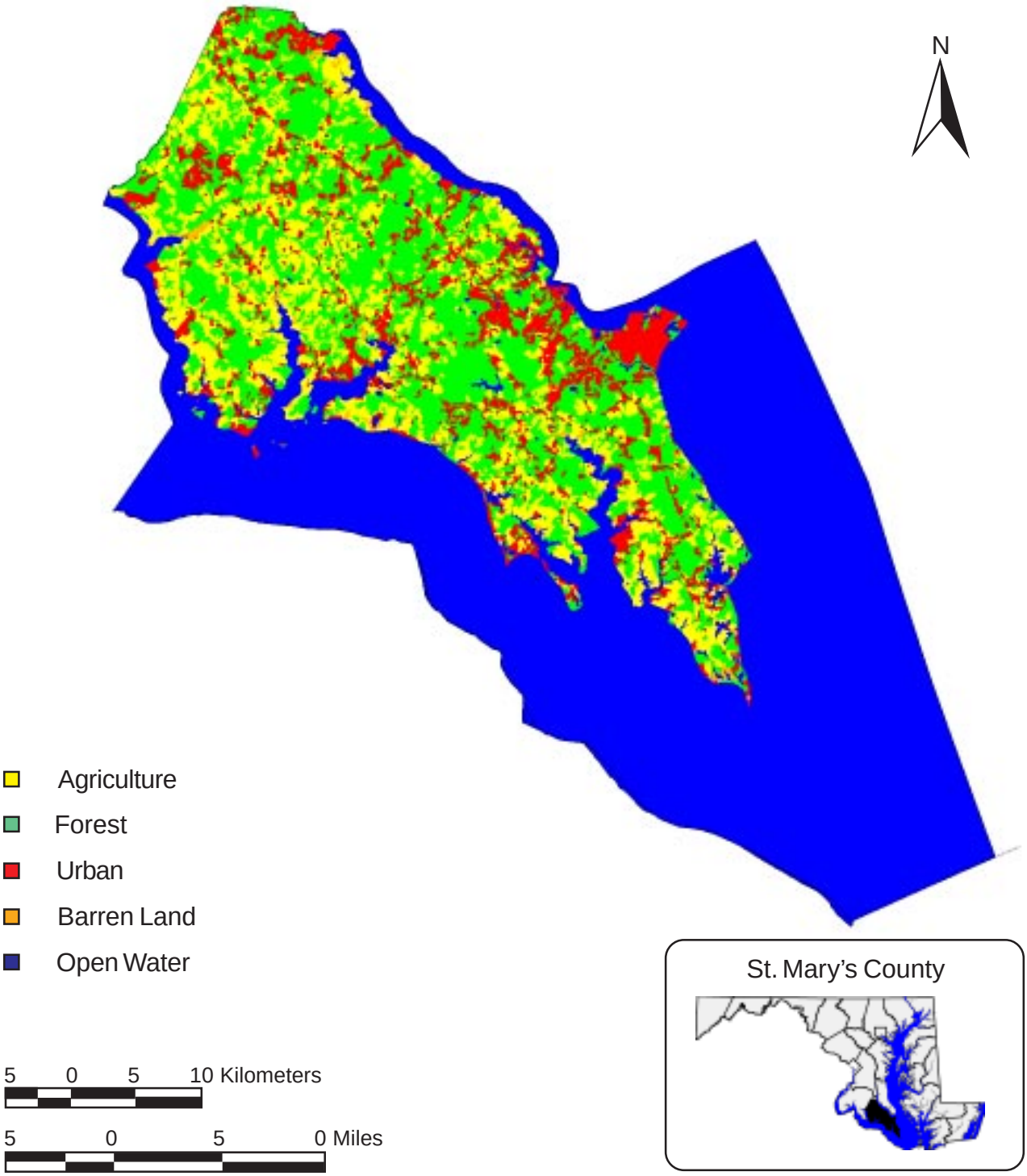


**Figure 1.** Land use in St. Mary's County (MOP 1994).

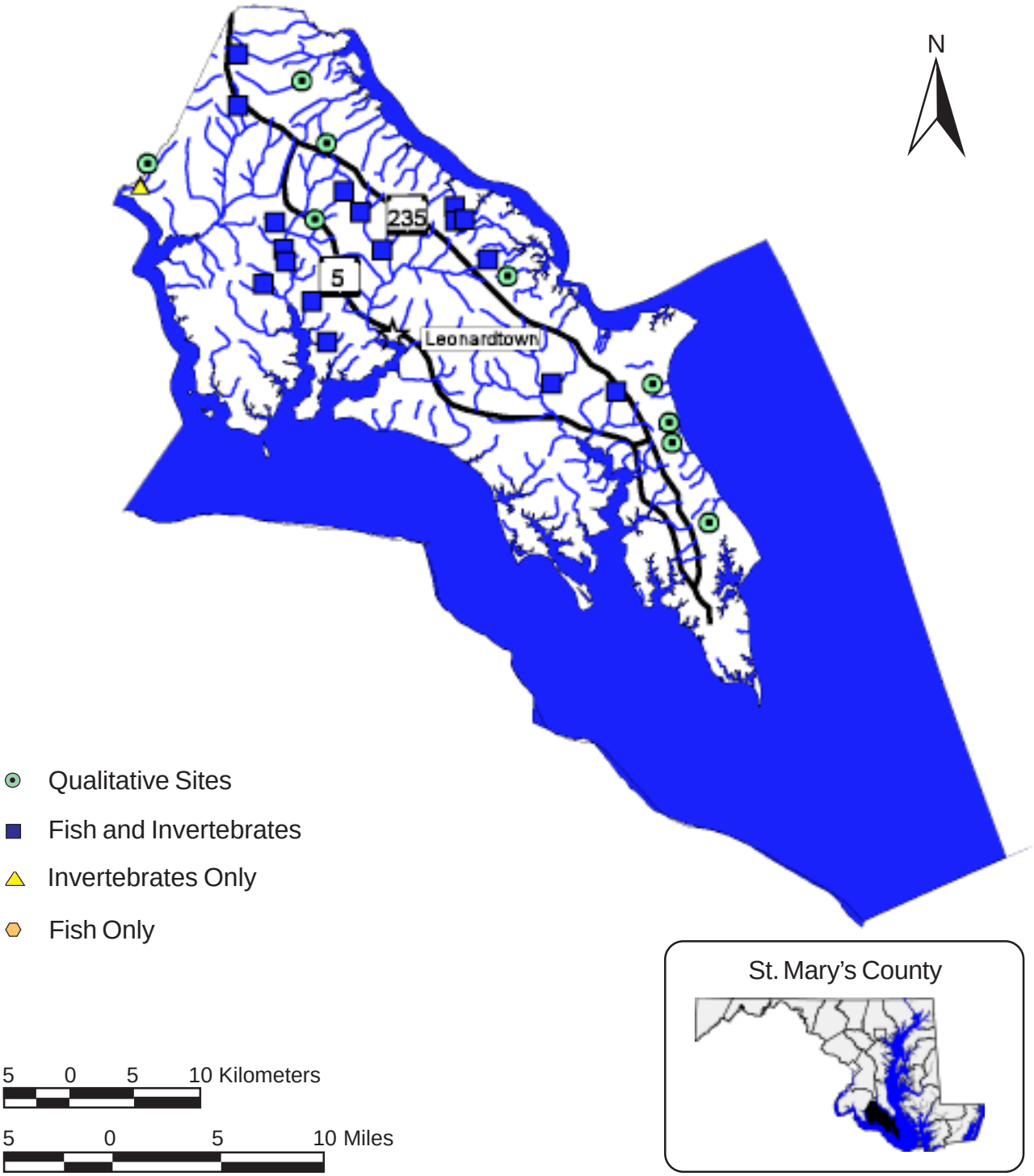


**Figure 2.** Location of Maryland Biological Stream Survey sites in St. Mary's County, 1994-1997.

**Table 2.** Percent occurrence of fish species collected at Maryland Biological Stream Survey sites in St. Mary's County, 1994-1997.

| Family | Common Name | Scientific Name | Number of Occurrences | Percent Occurrence |
|---|---|---|---|---|
| Petromyzontidae | least brook lamprey | *Lampetra aepyptera* | 10 | 58.82 |
| | sea lamprey | *Petromyzon marinus* | 4 | 23.53 |
| Anguillidae | American eel | *Anguilla rostrata* | 13 | 76.47 |
| Cyprinidae | rosyside dace | *Clinostomus funduloides* | 2 | 11.76 |
| | golden shiner | *Notemigonus crysoleucas* | 1 | 5.88 |
| | spottail shiner [1] | *Notropis hudsonius* | | |
| | swallowtail shiner | *Notropis procne* | 2 | 11.76 |
| | blacknose dace | *Rhinichthys atratulus* | 8 | 47.06 |
| | fallfish | *Semotilus corporalis* | 5 | 29.41 |
| Catostomidae | creek chubsucker | *Erimyzon oblongus* | 7 | 41.18 |
| Ictaluridae | tadpole madtom | *Noturus gyrinus* | 1 | 5.88 |
| | margined madtom | *Noturus insignis* | 2 | 11.76 |
| Esocidae | chain pickerel | *Esox niger* | 3 | 17.65 |
| Umbridae | eastern mudminnow | *Umbra pygmaea* | 16 | 94.12 |
| Aphredoderidae | pirate perch | *Aphredoderus sayanus* | 6 | 35.29 |
| Centrarchidae | redbreast sunfish | *Lepomis auritus* | 7 | 41.18 |
| | pumpkinseed | *Lepomis gibbosus* | 4 | 23.53 |
| | bluegill | *Lepomis machrochirus* | 4 | 23.53 |
| | largemouth bass | *Micropterus salmoides* | 3 | 17.65 |
| Percidae | tessellated darter | *Etheostoma olmstedi* | 9 | 52.94 |

[1] Qualitative Sites

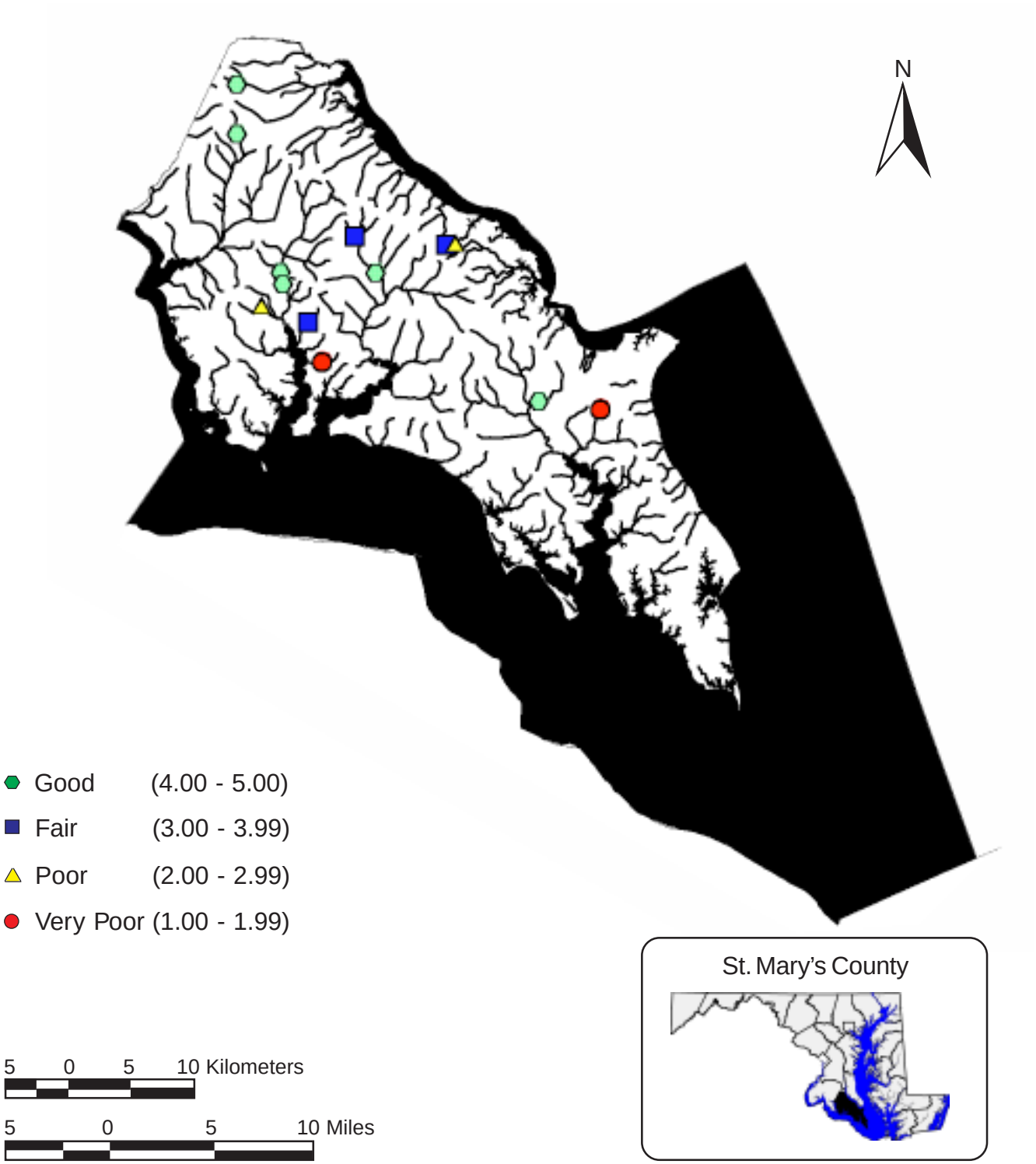


**Figure 3.** Stream ecological conditions based on the Fish Index of Biotic Integrity (F-IBI) at Maryland Biological Stream Survey sites in St. Mary's County, 1994-1997.

**Table 3.** Tolerance Value (TV)[1], Functional Feeding Group (FFG), Habit, and Percent Occurrence of benthic macroinvertebrate taxa[2] collected at Maryland Biological Stream Survey sites in St. Mary's County, 1994-1997. Abbreviations of habits are as follows: bu - burrower, cn - clinger, cb - climber, sp - sprawler, dv - diver, and sk - skater.

| Class | Order | Family | Genus | TV | FFG | Habit | Percent Occurrence |
|---|---|---|---|---|---|---|---|
| Gastropoda | Basommatophora | Lymnaeidae | *Pseudosuccinea* Sp. | 6 | Collector | cb | 25.56 |
| | | Physidae | *Physella* Sp. | 8 | Scraper | cb | 11.11 |
| | | Planorbidae | *Helisoma* Sp. | 6 | Scraper | cb | 5.56 |
| Pelecypoda | Veneroida | Corbiculidae | *Corbicula* Sp. | 6 | Filterer | bu | 5.56 |
| | | Sphaeriidae | *Pisidium* Sp. | 8 | Filterer | bu | 11.11 |
| | | | *Sphaerium* Sp. | 8 | Filterer | bu | 5.56 |
| Malacostraca | Amphipoda | Crangonyctidae | *Crangonyx* Sp. | 4 | Collector | sp | 38.89 |
| | | Gammaridae | *Gammarus* Sp. | 6 | Shredder | sp | 44.44 |
| Malacostraca | Isopoda | Asellidae | *Caecidotea* Sp. | 8 | Collector | sp | 44.44 |
| Insecta | Ephemeroptera | Ameletidae | *Ameletus* Sp. | 0 | Collector | sw, cb | 5.56 |
| | | Baetidae | *Acerpenna* Sp. | 4 | Collector | sw, cn | 50.00 |
| | | | *Centroptilum* Sp. | 2 | Collector | sw, cn | 5.56 |
| | | Caenidae | *Caenis* Sp. | 7 | Collector | sp | 5.56 |
| | | Ephemerellidae | *Ephemerella* Sp. | 2 | Collector | cn, sw | 50.00 |
| | | | *Eurylophella* Sp. | 4 | Scraper | cn, sp | 22.22 |
| | | Heptageniidae | *Stenonema* Sp. | 4 | Scraper | cn | 33.33 |
| | | Leptophlebiidae | | | Collector | sw, cn | 5.56 |
| | | | *Habrophlebia* Sp. | | Collector | sw, cn, sp | 5.56 |
| | | | *Leptophlebia* Sp. | 4 | Collector | sw, cn, sp | 22.22 |
| | | | *Paraleptophlebia* Sp. | 2 | Collector | sw, cn, sp | 5.56 |
| Insecta | Odonata | Aeshnidae | *Boyeria* Sp. | 2 | Predator | cb, sp | 11.11 |
| | | Calopterygidae | *Calopteryx* Sp. | 6 | Predator | cb | 11.11 |
| | | Cordulegastridae | *Cordulegaster* Sp. | 3 | Predator | bu | 22.22 |
| | | Corduliidae | *Macromia* Sp. | 3 | Predator | sp | 11.11 |
| | | | *Somatochlora* Sp. | 1 | Predator | sp | 5.56 |
| | | Gomphidae | *Arigomphus* Sp. | | Predator | bu | 5.56 |
| | | | *Gomphus* Sp. | 5 | Predator | bu | 11.11 |
| | | | *Stylogomphus* Sp. | | Predator | bu | 11.11 |
| Insecta | Plecoptera | Capniidae | *Paracapnia* Sp. | 1 | Shredder | - | 5.56 |
| | | Chloroperlidae | | | Predator | cn | 5.56 |
| | | | *Haploperla* Sp. | | Predator | cn | 11.11 |
| | | | *Sweltsa* Sp. | | Predator | cn | 16.67 |
| | | Leuctridae | *Leuctra* Sp. | 0 | Shredder | cn | 27.78 |
| | | Nemouridae | | | Shredder | sp, cn | 11.11 |
| | | | *Amphinemura* Sp. | 3 | Shredder | sp, cn | 44.44 |
| | | | *Ostrocerca* Sp. | | Shredder | sp, cn | 5.56 |
| | | | *Prostoia* Sp. | | Shredder | sp, cn | 5.56 |
| | | | *Soyedina* Sp. | | Shredder | sp, cn | 5.56 |
| | | Perlidae | | | Predator | cn | 5.56 |
| | | | *Eccoptura* Sp. | | Predator | cn | 22.22 |
| | | | *Perlesta* Sp. | 4 | Predator | cn | 5.56 |
| | | Perlodidae | *Isoperla* Sp. | 2 | Predator | cn, sp | 38.89 |
| | | Taeniopterygidae | *Strophopteryx* Sp. | | Shredder | sp, cn | 5.56 |
| Insecta | Hemiptera | Gerridae | *Trepobates* Sp. | | Predator | skater | 5.56 |
| | | Veliidae | *Microvelia* Sp. | 6 | Predator | skater | 11.11 |
| Insecta | Megaloptera | Corydalidae | *Nigronia* Sp. | 0 | Predator | cn, cb | 11.11 |
| | | Sialidae | *Sialis* Sp. | 4 | Predator | bu, cb, cn | 5.56 |

**Table 3 (cont.).** Tolerance Value (TV)[1], Functional Feeding Group (FFG), Habit, and Percent Occurrence of benthic macroinvertebrate taxa[2] collected at Maryland Biological Stream Survey sites in St. Mary's County, 1994-1997. Abbreviations of habits are as follows: bu - burrower, cn - clinger, cb - climber, sp - sprawler, dv - diver, and sk - skater.

| Class | Order | Family | Genus | TV | FFG | Habit | Percent Occurrence |
|---|---|---|---|---|---|---|---|
| Insecta | Trichoptera | Brachycentridae | *Brachycentrus* Sp. | 1 | Filterer | cn | 11.11 |
| | | Glossosomatidae | *Agapetus* Sp. | 2 | Scraper | cn | 5.56 |
| | | Hydropsychidae | *Cheumatopsyche* Sp. | 5 | Filterer | cn | 27.78 |
| | | | *Diplectrona* Sp. | 2 | Filterer | cn | 11.11 |
| | | | *Hydropsyche* Sp. | 6 | Filterer | cn | 22.22 |
| | | Lepidostomatidae | *Lepidostoma* Sp. | 3 | Shredder | cb, sp, cn | 11.11 |
| | | Leptoceridae | *Ceraclea* Sp. | 3 | Collector | sp, cb | 5.56 |
| | | | *Nectopsyche* Sp. | 3 | Shredder | cb, sw | 5.56 |
| | | | *Triaenodes* Sp. | 6 | Shredder | sw, cb | 5.56 |
| | | Limnephilidae | | | Shredder | cb, sp, cn | 11.11 |
| | | | *Ironoquia* Sp. | 3 | Shredder | sp | 5.56 |
| | | | *Pycnopsyche* Sp. | 4 | Shredder | sp, cb, cn | 55.56 |
| | | Philopotamidae | *Wormaldia* Sp. | | Filterer | cn | 5.56 |
| | | Phryganeidae | *Ptilostomis* Sp. | 5 | Shredder | cb | 16.67 |
| | | Polycentropodidae | *Polycentropus* Sp. | 5 | Filterer | cn | 11.11 |
| | | Psychomyiidae | *Lype* Sp. | 2 | Scraper | cn | 11.11 |
| | | Rhyacophilidae | *Rhyacophila* Sp. | 1 | Predator | cn | 5.56 |
| | | Uenoidae | *Neophylax* Sp. | 3 | Scraper | cn | 38.89 |
| Insecta | Lepidoptera | | | 6 | | | 5.56 |
| Insecta | Coleoptera | Dryopidae | *Helichus* Sp. | 5 | Scraper | cn | 5.56 |
| | | Dytiscidae | | 5 | Predator | sw, dv | 5.56 |
| | | | *Hydroporus* Sp. | 5 | Predator | sw, cb | 5.56 |
| | | Elmidae | *Ancyronyx* Sp. | 2 | Scraper | cn, sp | 11.11 |
| | | | *Dubiraphia* Sp. | 6 | Scraper | cn, cb | 16.67 |
| | | | *Macronychus* Sp. | 4 | Scraper | cn | 22.22 |
| | | | *Optioservus* Sp. | 4 | Scraper | cn | 11.11 |
| | | | *Oulimnius* Sp. | 2 | Scraper | cn | 38.89 |
| | | | *Stenelmis* Sp. | 6 | Scraper | cn | 16.67 |
| | | Gyrinidae | *Dineutus* Sp. | 4 | Predator | sw, dv | 5.56 |
| | | Hydrophilidae | *Hydrobius* Sp. | 5 | Collector | cb, cn, sp | 5.56 |
| | | Psephenidae | *Psephenus* Sp. | 4 | Scraper | cn | 11.11 |
| | | Ptilodactylidae | *Anchytarsus* Sp. | 4 | Shredder | cn | 11.11 |
| Insecta | Diptera | Ceratopogonidae | *Bezzia* Sp. | 6 | Predator | bu | 11.11 |
| | | | *Culicoides* Sp. | 10 | Predator | bu | 5.56 |
| | | | *Probezzia* Sp. | 6 | Predator | bu | 5.56 |
| | | Chironomidae | *Ablabesmyia* Sp. | 8 | Predator | sp | 11.11 |
| | | | *Brillia* Sp. | 5 | Shredder | bu, sp | 5.56 |
| | | | *Conchapelopia* Sp. | 6 | Predator | sp | 44.44 |
| | | | *Corynoneura* Sp. | 7 | Collector | sp | 22.22 |
| | | | *Cricotopus* Sp. | 7 | Shredder | cn, bu | 5.56 |
| | | | *Cricotopus/ Orthocladius* Sp. | | Shredder | | 5.56 |
| | | | *Eukiefferiella* Sp. | 8 | Collector | sp | 5.56 |
| | | | *Heleniella* Sp. | | Predator | sp | 5.56 |
| | | | *Hydrobaenus* Sp. | 8 | Scraper | sp | 11.11 |
| | | | *Labrundinia* Sp. | 7 | Predator | sp | 5.56 |
| | | | *Micropsectra* Sp. | 7 | Collector | cb, sp | 22.22 |

**Table 3 (cont.).** Tolerance Value (TV)[1], Functional Feeding Group (FFG), Habit, and Percent Occurrence of benthic macroinvertebrate taxa[2] collected at Maryland Biological Stream Survey sites in St. Mary's County, 1994-1997. Abbreviations of habits are as follows: bu - burrower, cn - clinger, cb - climber, sp - sprawler, dv - diver, and sk - skater.

| Class | Order | Family | Genus | TV | FFG | Habit | Percent Occurrence |
|---|---|---|---|---|---|---|---|
| | | | *Microtendipes* Sp. | 6 | Filterer | cn | 11.11 |
| | | | *Nanocladius* Sp. | 3 | Collector | sp | 5.56 |
| | | | *Orthocladius* Sp. | 6 | Collector | sp, bu | 22.22 |
| | | | *Parametriocnemus* Sp. | 5 | Collector | sp | 66.67 |
| | | | *Paratanytarsus* Sp. | 6 | Collector | sp | 5.56 |
| | | | *Paratrichocladius* Sp. | | Collector | sp | 5.56 |
| | | | *Pentaneura* Sp. | 6 | Predator | sp | 5.56 |
| | | | *Polypedilum* Sp. | 6 | Shredder | cb, cn | 27.78 |
| | | | *Potthastia* Sp. | 2 | Collector | sp | 5.56 |
| | | | *Procladius* Sp. | 9 | Predator | sp | 5.56 |
| | | | *Rheocricotopus* Sp. | 6 | Collector | sp | 11.11 |
| | | | *Rheotanytarsus* Sp. | 6 | Filterer | cn | 16.67 |
| | | | *Stictochironomus* Sp. | 9 | Collector | bu | 5.56 |
| | | | *Symposiocladius* Sp. | | Predator | sp | 5.56 |
| | | | *Sympotthastia* Sp. | 2 | Collector | sp | 5.56 |
| | | | *Tanytarsus* Sp. | 6 | Filterer | cb, cn | 11.11 |
| | | | *Thienemanniella* Sp. | 6 | Collector | sp | 11.11 |
| | | | *Thienemannimyia* Sp. | | Predator | sp | 11.11 |
| | | | *Trissopelopia* Sp. | | Predator | sp | 5.56 |
| | | | *Tvetenia* Sp. | 5 | Collector | sp | 33.33 |
| | | | CHIRONOMINAE | 6 | Collector | | 5.56 |
| | | | *Zavrelia* Sp. | 4 | Collector | cb, sp, cn | 5.56 |
| | | | *Zavrelimyia* Sp. | 8 | Predator | sp | 22.22 |
| | | Dixidae | *Dixa* Sp. | 4 | Predator | sw, cb | 5.56 |
| | | Empididae | *Chelifera* Sp. | | Predator | sp, bu | 5.56 |
| | | | *Hemerodromia* Sp. | 6 | Predator | sp, bu | 16.67 |
| | | Muscidae | | 7 | Predator | sp | 5.56 |
| | | Simuliidae | *Prosimulium* Sp. | 7 | Filterer | cn | 44.44 |
| | | | *Simulium* Sp. | 7 | Filterer | cn | 27.78 |
| | | | *Stegopterna* Sp. | 7 | Filterer | cn | 50.00 |
| | | Tabanidae | *Chrysops* Sp. | 7 | Predator | sp, bu | 11.11 |
| | | | *Tabanus* Sp. | 5 | Predator | sp, bu | 11.11 |
| | | Tipulidae | *Dicranota* Sp. | 4 | Predator | sp, bu | 5.56 |
| | | | *Hexatoma* Sp. | 4 | Predator | bu, sp | 38.89 |
| | | | *Limonia* Sp. | 6 | Shredder | bu, sp | 5.56 |
| | | | *Ormosia* Sp. | | Collector | bu | 5.56 |
| | | | *Pseudolimnophila* Sp. | 2 | Predator | bu | 11.11 |
| | | | *Tipula* Sp. | 4 | Shredder | bu | 33.33 |

[1] Tolerance values are on a 0 (extremely sensitive) to 10 (tolerant) scale.

[2] Taxa not identified to genus are presented in capital letters. Subfamily - Chironominae.

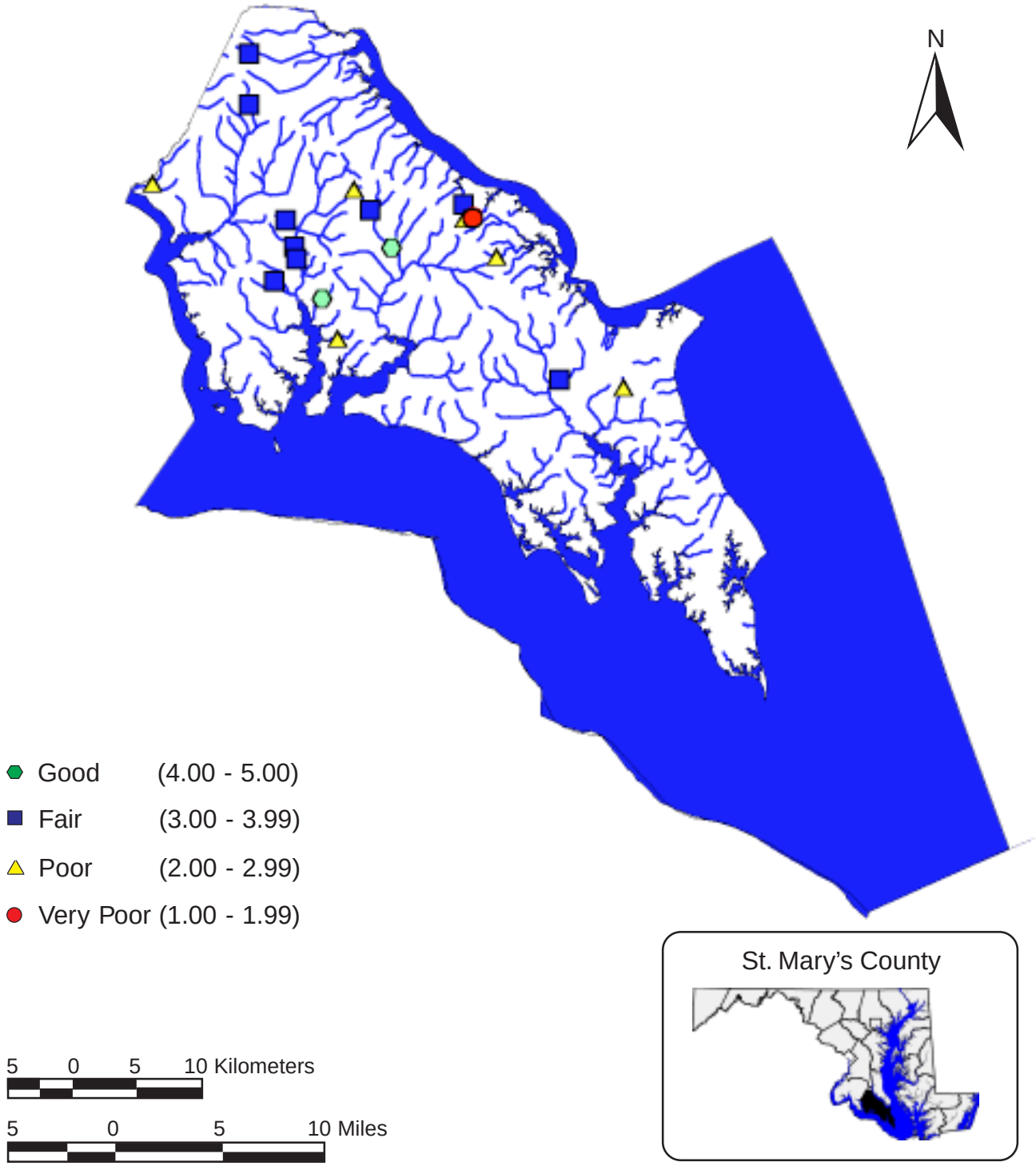


**Figure 4.** Stream ecological conditions based on the Benthic Macroinvertebrate Index of Biotic Integrity (B-IBI) at Maryland Biological Stream Survey sites in St. Mary's County, 1994-1997.

**Table 4.** Percent occurrence of reptile and amphibian species collected at Maryland Biological Stream Survey sites in St. Mary's County, 1994-1997.

| Family | Common Name | Scientific Name | Number of Occurrences | Percent Occurrence |
|---|---|---|---|---|
| Plethodontidae | eastern mud salamander | *Pseudotriton m. montanus* | 1 | 5.88 |
| | northern two-lined salamander | *Eurycea bislineata* | 8 | 47.06 |
| | red salamander | *Pseudotriton ruber* | 1 | 5.88 |
| Bufonidae | Fowler's toad | *Bufo woodhousii fowleri* | 2 | 11.76 |
| Ranidae | bullfrog | *Rana catesbeiana* | 10 | 58.82 |
| | green frog | *Rana clamitans melanota* | 13 | 76.47 |
| | pickerel frog | *Rana palaustris* | 11 | 64.71 |
| | southern leopard frog | *Rana utricularia* | 9 | 52.94 |
| Chelydridae | common snapping turtle | *Chelydra serpentina* | 3 | 17.65 |
| Emydidae | eastern box turtle | *Terrapene c. carolina* | 5 | 29.41 |
| Colubridae | black rat snake | *Elaphe o. obsoleta* | 1 | 5.88 |
| | eastern worm snake | *Carphophis a. amoenus* | 1 | 5.88 |
| | northern black racer | *Coluber c. constrictor* | 2 | 11.76 |
| | northern ringneck snake | *Diadophis punctatus edwardsii* | 1 | 5.88 |
| | northern water snake | *Nerodia s. sipedon* | 1 | 5.88 |

**Table 5.** Physical habitat data for Maryland Biological Stream Survey sites in St. Mary's County, 1994-1997.

| Site | Instream Habitat[1] | Velocity/Depth Diversity[1] | | Riffle Quality[1] | | Percent Shading[1] | | Number of Woody Debris | | Percent Channel Flow[1] | | Bank Stability[1] | | Aesthetic Rating[1] |
|---|---|---|---|---|---|---|---|---|---|---|---|---|---|---|
| | | Epifaunal Substrate[1] | | Pool Quality[1] | | Percent Embeddedness[1] | | Maximum Depth (cm)[1] | | Number of Rootwads | | Channel Alteration[1] | | Riparian Width (m)[1] |
| SM-S-006-212-95 | 17 | 17 | 15 | 15 | 15 | 10 | 60 | 85 | 13 | 3 | 70 | 11 | 11 | 50 | 15 |
| SM-S-007-138-95 | 5 | 4 | 2 | 1 | 6 | 100 | 95 | 23 | 2 | 3 | 100 | 15 | 7 | 50 | 17 |
| SM-S-036-107-95 | 7 | 6 | 8 | 6 | 10 | 35 | 90 | 32 | 0 | 1 | 35 | 7 | 10 | 50 | 18 |
| SM-S-039-127-95 | 12 | 14 | 13 | 13 | 8 | 20 | 95 | 50 | 10 | 2 | 55 | 10 | 8 | 50 | 17 |
| SM-S-040-128-95 | 8 | 5 | 8 | 7 | 7 | 60 | 95 | 32 | 9 | 1 | 40 | 6 | 4 | 50 | 7 |
| SM-S-051-132-95 | 12 | 11 | 12 | 11 | 9 | 5 | 90 | 79 | 8 | 3 | 50 | 8 | 5 | 50 | 19 |
| SM-S-090-238-97 | 14 | 15 | 13 | 14 | 16 | 15 | 95 | 55 | 2 | 5 | 90 | 10 | 11 | 12 | 18 |
| SM-S-104-126-95 | 5 | 3 | 6 | 6 | 6 | 95 | 90 | 38 | 5 | 2 | 50 | 5 | 7 | 50 | 11 |
| SM-S-111-112-95 | 16 | 13 | 15 | 16 | 15 | 75 | 85 | 70 | 5 | 4 | 60 | 9 | 9 | 50 | 17 |
| SM-S-125-142-97 | 7 | 11 | 7 | 6 | 11 | 40 | 90 | 42 | 0 | 2 | 90 | 9 | 9 | 50 | 18 |
| SM-S-166-143-97 | 4 | 12 | 7 | 6 | 11 | 54 | 65 | 14 | 1 | 3 | 88 | 16 | 16 | 50 | 18 |
| SM-S-199-302-95 | 15 | 13 | 11 | 18 | 7 | 60 | 95 | 101 | 10 | 4 | 80 | 6 | 5 | 50 | 15 |
| SM-S-209-105-95 | 3 | 1 | 6 | 8 | 2 | 100 | 95 | 76 | 1 | 1 | 60 | 4 | 5 | 15 | 10 |
| SM-S-214-1-94 | 6 | 2 | 8 | 7 | 7 | 100 | 98 | 44 | 7 | | 80 | 4 | 6 | 50 | 16 |
| SM-S-214-2-94 | 3 | 1 | 11 | 6 | 7 | 97 | 75 | 44 | 6 | | 90 | 6 | 6 | 50 | 15 |
| SM-S-237-103-95 | 6 | 6 | 8 | 6 | 10 | 65 | 90 | 41 | 5 | 2 | 85 | 9 | 10 | 6 | 18 |
| SM-S-239-310-95 | 16 | 11 | 14 | 19 | 16 | 60 | 70 | 109 | 13 | 5 | 90 | 8 | 9 | 50 | 17 |

[1] MBSS Qualitative Habitat Metric - See Appendix B for Guidance

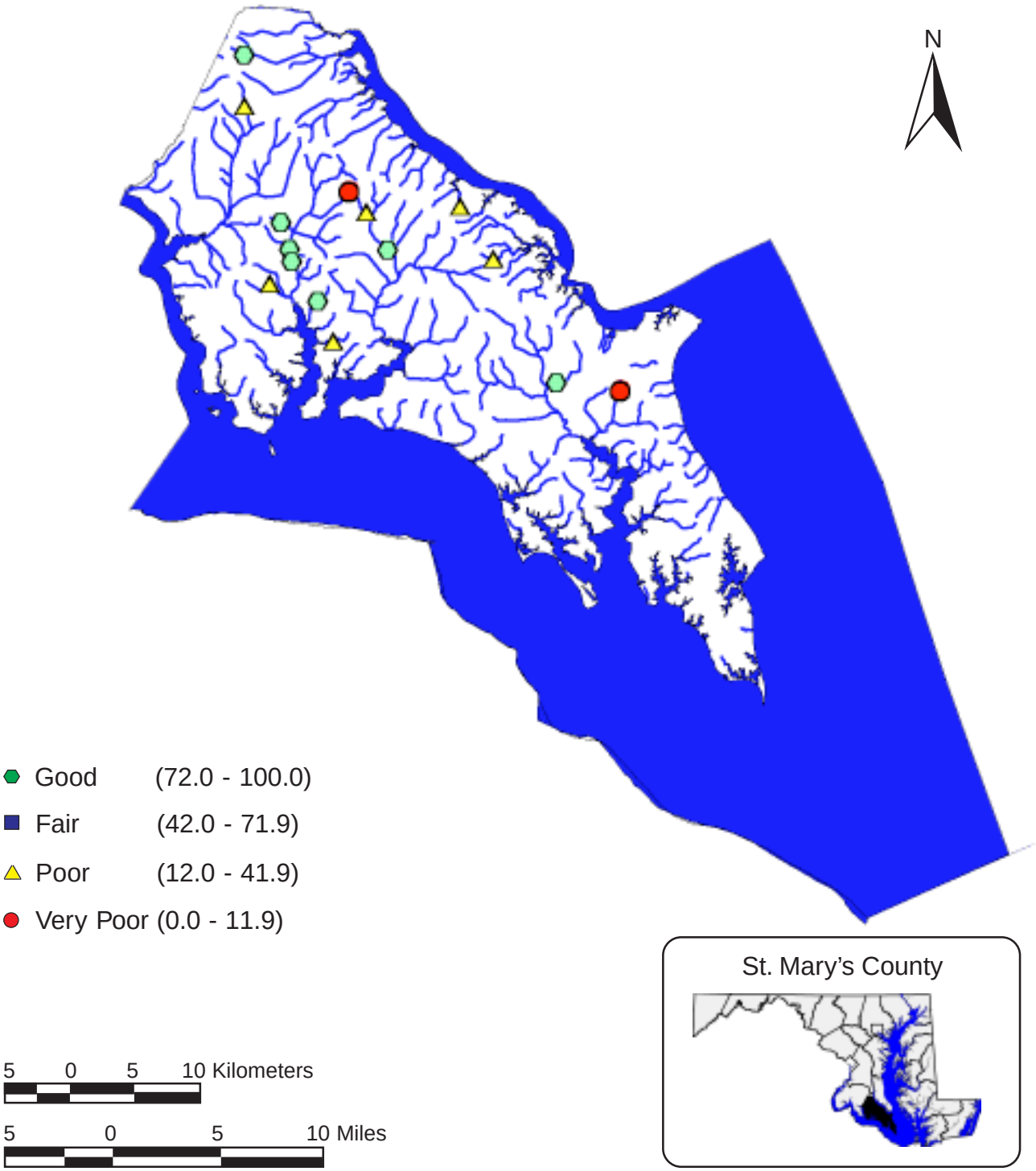


**Figure 5.** Stream ecological conditions based on the Physical Habitat Index (PHI) at Maryland Biological Stream Survey sites in St. Mary's County, 1994-1997.

**Table 6.** Fish Index of Biotic Integrity (F-IBI), Benthic Macroinvertebrate Index of Biotic Integrity (B-IBI), and Physical Habitat Index (PHI) scores at Maryland Biological Stream Survey sites in St. Mary's County, 1994-1997.

| Site | Stream Name | F-IBI | B-IBI | PHI |
|---|---|---|---|---|
| SM-S-006-212-95 | Burnt Mill Creek | 4.00 | 4.4 | 94.59 |
| SM-S-007-138-95 | Penbrook Run | 1.00 | 2.4 | 4.74 |
| SM-S-036-107-95 | Dynard Run | 2.50 | 3.6 | 33.24 |
| SM-S-039-127-95 | Ut St Clements Bay | 3.75 | 4.7 | 81.52 |
| SM-S-040-128-95 | Ut Burnt Mill Creek | 3.75 | 3.9 | 20.45 |
| SM-S-051-132-95 | Ut St Clements Creek | | 3.9 | 85.18 |
| SM-S-090-238-97 | Un Trib To Killpeck Cr | 4.50 | 3.3 | 87.87 |
| SM-S-104-126-95 | Ut St Clements Creek | | 2.4 | 10.73 |
| SM-S-111-112-95 | Jarboesville Run | 4.50 | 3.9 | 88.99 |
| SM-S-116-214-95 | Budds Creek | | 2.1 | |
| SM-S-125-142-97 | Cuckold Cr | | 2.7 | 32.03 |
| SM-S-166-143-97 | Un Trib To Coles Cr | | 3.0 | 17.57 |
| SM-S-199-302-95 | St Clements Creek | 4.25 | 3.9 | 89.62 |
| SM-S-209-105-95 | Cecil Creek | 1.00 | 2.7 | 14.06 |
| SM-S-214-1-94 | St Thomas Cr | 3.25 | 2.1 | |
| SM-S-214-2-94 | St Thomas Cr | 2.50 | 1.6 | |
| SM-S-237-103-95 | Chaptico Creek | 4.50 | 3.6 | 26.13 |
| SM-S-239-310-95 | St Clements Creek | 4.00 | 3.6 | 94.81 |

**Table 7.** Water chemistry data collected at Maryland Biological Stream Survey sites in St. Mary's County, 1994-1997.

| Site | pH | Conductivity (µS/cm) | Acid Neutralizing Capacity (µeq/L) | Nitrate (mg/L) | Sulfate (mg/L) | Dissolved Oxygen (mg/L) | Dissolved Organic Carbon (mg/L) |
|---|---|---|---|---|---|---|---|
| SM-S-006-212-95 | 6.92 | 0.089 | 364.41 | 0.451 | 8.382 | 6.50 | 8.00 |
| SM-S-007-138-95 | 5.44 | 0.040 | 50.89 | 0.257 | 4.258 | 2.20 | 7.00 |
| SM-S-036-107-95 | 7.25 | 0.118 | 611.43 | 0.423 | 12.451 | 7.40 | 4.00 |
| SM-S-039-127-95 | 7.24 | 0.136 | 596.28 | 1.375 | 16.212 | 7.30 | 3.00 |
| SM-S-040-128-95 | 6.32 | 0.060 | 113.56 | 0.223 | 8.342 | 6.50 | 6.00 |
| SM-S-051-132-95 | 6.86 | 0.062 | 313.88 | 0.200 | 7.048 | 7.50 | 4.00 |
| SM-S-090-238-97 | 7.08 | 0.175 | 275.20 | 3.009 | 18.051 | 10.00 | 2.60 |
| SM-S-104-126-95 | 5.85 | 0.077 | 69.82 | 0.293 | 8.914 | 5.00 | 6.00 |
| SM-S-111-112-95 | 6.51 | 0.068 | 182.65 | 0.353 | 4.642 | 6.90 | 7.00 |
| SM-S-116-214-95 | 6.28 | 0.071 | 72.17 | | 17.093 | | 4.00 |
| SM-S-125-142-97 | 6.54 | 0.090 | 159.60 | 0.573 | 12.310 | 9.90 | 1.40 |
| SM-S-166-143-97 | 7.85 | 0.275 | 2456.60 | 0.323 | 20.070 | 9.20 | 2.00 |
| SM-S-199-302-95 | 7.10 | 0.103 | 452.86 | 0.460 | 10.310 | 4.70 | 6.00 |
| SM-S-209-105-95 | 6.19 | 0.074 | 107.35 | 0.194 | 16.160 | 3.50 | 6.00 |
| SM-S-214-1-94 | | | | | | | |
| SM-S-214-2-94 | 7.46 | 0.185 | 1490.73 | 0.087 | 16.531 | | 2.00 |
| SM-S-237-103-95 | 6.94 | 0.120 | 499.16 | 1.383 | 16.336 | 6.30 | 6.00 |
| SM-S-239-310-95 | 7.07 | 0.096 | 457.95 | 0.492 | 10.639 | 7.30 | 6.00 |

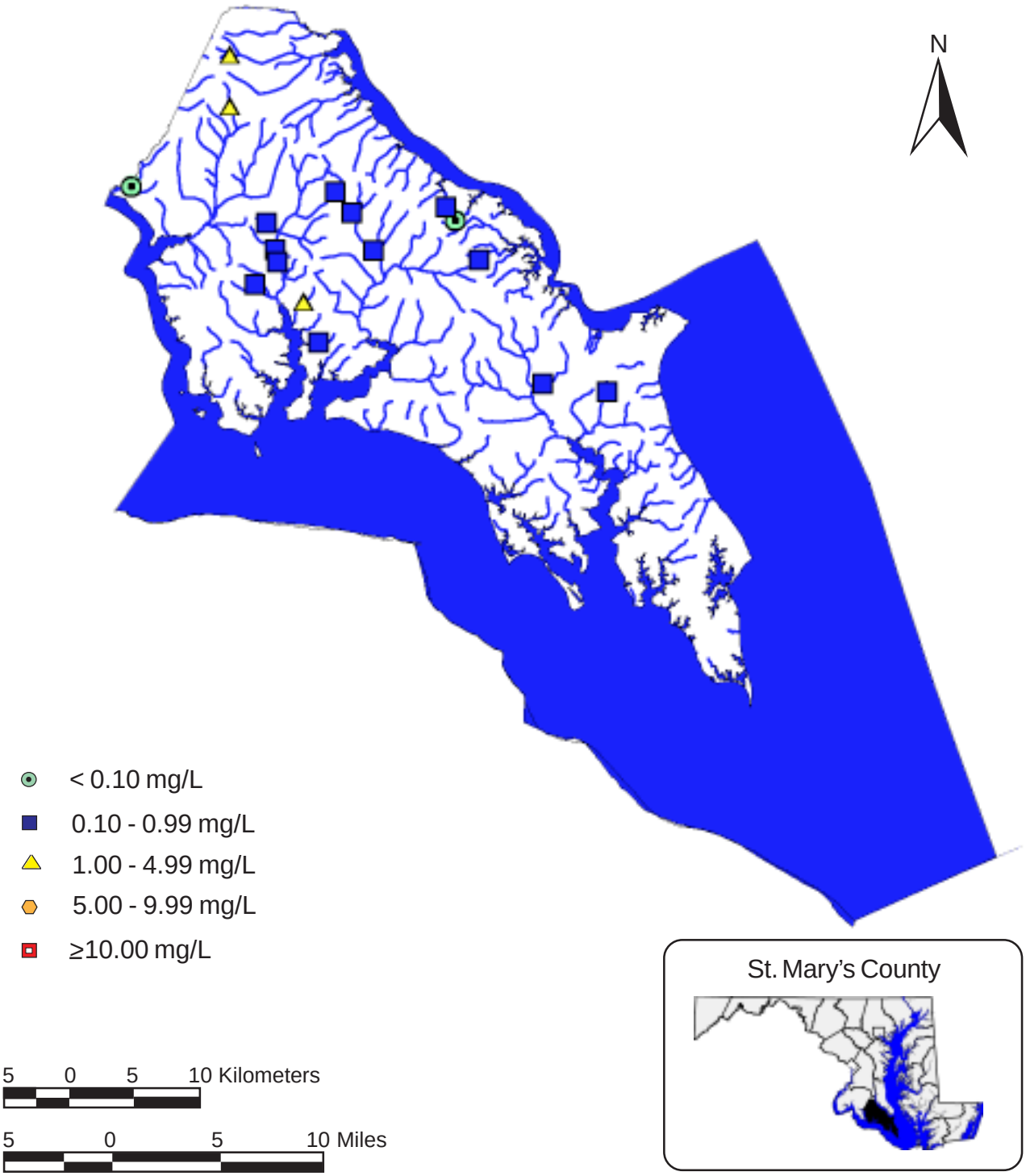


**Figure 6.** Nitrate-nitrogen concentrations (mg/L) at Maryland Biological Stream Survey sites in St. Mary's County, 1994-1997.

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**Appendix A.** Summary of the types of data collected at Maryland Biological Stream Survey sites in St. Maryl's County, 1994-1997. Abbreviations used are as follows: F-IBI - Fish Index of Biotic Integrity; B-IBI Benthic Macroinvertebrate Index of Biotic Integrity; Fam.IBI - Family-Level Benthic Macroinvertebrate Index of Biotic Integrity; PHI - Physical Habitat Index.

| | | | Benthic Macroinvertebrate | | Habitat | | F-IBI | Fam. IBI | |
|---|---|---|---|---|---|---|---|---|---|
| **Site** | **Stream Name** | **Fish** | | **Herpetofauna** | | **Water Chemistry** | | **B-IBI** | **PHI** |
| M-S-006-212-95 | Burnt Mill Creek | X | X | X | X | X | X | X | X |
| SM-S-007-138-95 | Penbrook Run | X | X | X | X | X | X | X | X |
| SM-S-036-107-95 | Dynard Run | X | X | X | X | X | X | X | X |
| SM-S-039-127-95 | Ut St Clements Bay | X | X | X | X | X | X | X | X |
| SM-S-040-128-95 | Ut Burnt Mill Creek | X | X | X | X | X | X | X | X |
| SM-S-051-132-95 | Ut St Clements Creek | X | X | X | X | X | | X | X |
| SM-S-090-238-97 | Un Trib To Killpeck Cr | X | X | X | X | X | X | X | X |
| SM-S-104-126-95 | Ut St Clements Creek | X | X | X | X | X | | X | X |
| SM-S-111-112-95 | Jarboesville Run | X | X | X | X | X | X | X | X |
| SM-S-116-214-95 | Budds Creek | | X | | | X | | X | |
| SM-S-125-142-97 | Cuckold Cr | X | X | X | X | X | | X | X |
| SM-S-166-143-97 | Un Trib To Coles Cr | X | X | X | X | X | | X | X |
| SM-S-199-302-95 | St Clements Creek | X | X | X | X | X | X | X | X |
| SM-S-209-105-95 | Cecil Creek | X | X | X | X | X | X | X | X |
| SM-S-214-1-94 | St Thomas Cr | X | X | X | X | | X | X | |
| SM-S-214-2-94 | St Thomas Cr | X | X | X | X | X | X | X | |
| SM-S-237-103-95 | Chaptico Creek | X | X | X | X | X | X | X | X |
| SM-S-239-310-95 | St Clements Creek | X | X | X | X | X | X | X | X |

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**Appendix B.** Physical habitat condition measured by the Maryland Biological Stream Survey, 1994-1997. All variables rated on a scale of 0 (poor) to 20 (optimal) unless otherwise noted.

## *SUBSTRATE AND INSTREAM COVER*

**Instream Habitat** is rated according to the perceived value of habitat to the fish community. Higher scores are assigned to sites with a variety of habitat types and particle sizes. In addition, higher scores are assigned to sites with a high degree of uneven substrate, including logs and rootwads. In streams where substrate types are favorable but flows are so low that fish are essentially precluded from using the habitat, low scores are assigned. If none of the habitat within a segment is useable by fish, a score of zero is assigned.

**Epifaunal Substrate** is rated based on the amount and variety of hard, stable substrates usable by benthic macroinvertebrates. Because they inhibit colonization, flocculent materials or fine sediments surrounding otherwise good substrates are assigned low scores. Scores are also reduced when substrates are less stable.

**Velocity/Depth Diversity** is rated based on the variety of velocity/depth regimes present at a site (slow-shallow, slow-deep, fast-shallow, and fast-deep). As with embeddedness, this metric varies by stream gradient.

**Pool/Glide/Eddy Quality** is rated based on the variety and spatial complexity of slow or still water habitat within the sample segment. In high-gradient streams, functionally important slow water habitat may exist in the form of larger eddies. Within a category, higher scores are assigned to segments which have undercut banks, woody debris or other types of cover for fish.

**Riffle/Run Quality** is based on the depth, complexity, and functional importance of riffle/run habitat in the segment, with highest scores assigned to segments dominated by deeper riffle/run areas, stable substrates, and a variety of current velocities.

**Embeddedness** is a percentage of surface area of larger particles that is surrounded by fine sediments on the stream bottom. In low gradient streams, embeddedness may be high even in relatively unimpaired watersheds.

## *CHANNEL CHARACTER*

**Channel Alteration** is a measure of large-scale changes in the shape of the stream channel. Channel alteration includes: concrete channels, artificial embankments, obvious straightening of the natural channel, rip-rap, or other structures, as well as recent bar development. Ratings for this metric are based on the presence of artificial structures as well as the existence, extent, and coarseness of point bars, side bars, and mid-channel bars which indicate the degree of flow fluctuations and substrate stability. Evidence of channelization may sometimes be seen in the form of berms that parallel the stream channel.

**Bank Stability** is rated based on the presence/absence of riparian vegetation and other stabilizing bank materials such as boulders and rootwads, and frequency/size of erosional areas. Sites with steep slopes are not penalized if banks are composed solely of stable materials.

**Channel Flow Status** is the percentage of the stream channel that has water, with subtractions made for exposed substrates and dewatered areas.

## *RIPARIAN CORRIDOR*

**Shading** is rated based on estimates of the degree and duration of shading at a site during summer, including any effects of shading caused by land forms.

**Appendix B (cont.).** Physical habitat condition measured by the Maryland Biological Stream Survey, 1994-1997. All variables rated on a scale of 0 (poor) to 20 (optimal) unless otherwise noted.

**Riparian Buffer** is rated according to the size and type of the vegetated riparian buffer zone at the site. Cultivated fields for agriculture that have bare soil to any extent are not considered as riparian buffers. At sites where the buffer width is variable, or direct delivery of storm runoff or sediment to the stream is evident or highly likely, the narrowest representative buffer width in the segment (e.g., 0 if parking lot runoff enters directly to the stream) is measured and recorded even though some of the stream segment may have a well developed riparian buffer.

## *AESTHETICS/REMOTENESS*

**Aesthetics** are rated according to the visual appeal of the site and presence/absence of human refuse, with highest scores assigned to stream segments with no human refuse and visually outstanding character.

**Remoteness** is rated based on the absence of detectable human activity and difficulty in accessing the segment.